

1 Distribution of Fragrances and PAHs in the Surface  
2 Seawater of the Sicily Channel, Central Mediterranean.

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20 **Keywords:** Fragrance Materials (FMs); Personal Care Products (PCPs); Polycyclic Aromatic  
21 Hydrocarbons (PAHs); Seawater; Mediterranean.

22

### 23 **Highlights**

24 First study on PCPs in Mediterranean open seawater.

25 Detailed distribution of PAHs in the Sicily Channel.

26 Salicylates result the most abundant FMs.

27 Pyrogenic sources of PAHs.

28 Marine currents and atmospheric transport are major drivers.

29

### 30 **Abstract**

31 The Mediterranean Sea is highly influenced by several anthropic pressures, including different kinds  
32 of organic pollutants. Fragrance Materials (FMs) and Polycyclic Aromatic Hydrocarbons (PAHs) were  
33 investigated in the surface seawater of the Sicily Channel in offshore and coastal areas. Total  
34 concentrations of FMs and PAHs resulted respectively up to 112 ng L<sup>-1</sup> and 43 ng L<sup>-1</sup>, with similar  
35 distributions of both classes of analytes: lower values were detected in some coastal samples, due  
36 to the upwelling of deep and unpolluted waters, while the presence of gyres probably accumulates  
37 contaminants in offshore areas. Confirming previous works, the allergenic and oestrogenic  
38 Salicylates generally resulted the most abundant FMs and diagnostic ratios indicated combustion  
39 processes as the sources of PAHs. The coupling of the well-known PAHs with a new class of  
40 Personal Care Products (PCPs) helped the identification of the major environmental drivers: the  
41 results highlighted the role of mesoscale hydrodynamics and suggested long-range atmospheric  
42 transport as key factors. The first detection of the selected FMs in open sea areas support the  
43 hypothesis of their environmental persistence.

44

## 45 **1. Introduction**

46 The Mediterranean Sea is a semi-enclosed basin surrounded by highly populated regions (Berrojalbiz  
47 et al., 2011). Anthropogenic pressures largely influence the environment, deriving from industrialization,  
48 urbanization and agriculture, as well as fishing, maritime traffic, harbor activities and tourism  
49 (Castro-Jiménez et al., 2012; Er-Raioui et al., 2009). The rapid increase of the demographic pressure  
50 and of the economic evolution lead to a large number of organic pollutant sources constituting a  
51 threat to the environment and biodiversity of the area (Castro-Jiménez et al., 2012). The  
52 Mediterranean Sea is therefore a region of special interest to study the factors controlling  
53 distribution, transport and fate of pollutants in marine waters (Berrojalbiz et al., 2011). However  
54 there are major knowledge gaps, since, in comparison to offshore waters, most studies focus on  
55 sediments and on local coastal areas influenced by proximity to inputs and runoff from land-based  
56 sources (Berrojalbiz et al., 2011; Giuliani et al., 2015). In the open sea other key factors, such as  
57 atmospheric transport and deposition, become important drivers as significant diffuse sources  
58 (Tsapakis et al., 2006).

59 Among organic pollutants, very little is known about the presence of Personal Care Products (PCPs)  
60 in the Mediterranean: few studies report the presence of PCPs in surface waters, Wastewater  
61 Treatment Plants (WWTPs), sediments and organisms (Arpin-Pont et al., 2016) and this information  
62 is limited to coastal marine environments. Fragrance Materials (FMs) are widespread used  
63 ingredients of PCPs and previously synthetic musks represent a category of contaminants whose  
64 environmental fate was mainly investigated by the literature (Daughton and Ternes, 1999). However,  
65 little is known about other different FMs and in this study we focus on 17 long-lasting and stable  
66 fragrances that are commercially available (Vecchiato et al., 2016). The trade names (Givaudan®) of  
67 the selected FMs are: Amberketal, Ambrofix, Amyl Salicylate, Benzyl Salicylate, Bourgeonal, Dupical,

68 Hexyl Salicylate, Isobutavan, Lemonile, Mefranal, Myraldene, Okoumal, Oranger Crystals, Pelargene,  
69 Peonile, Tridecene-2-Nitrile, Ultravanil (Table SI1). The common features of these FMs are a long  
70 persistence as fragrances (tenacity on blotter) and a chemical stability that allows their application in  
71 very different commercial products ([eindex.givaudan.com](http://eindex.givaudan.com)). Such characteristics are possible  
72 indications of their persistence in the environment. The selected FMs were recently found as  
73 contaminants in the Venice Lagoon, where sewage outlets largely emit these compounds into the  
74 surface seawater, with concentrations up to  $10 \mu\text{g L}^{-1}$  in the innermost urban canals (Vecchiato et al.,  
75 2016). Traces were found also in the coastal surface seawater of Terra Nova Bay, Antarctica, where  
76 emissions from the nearby research station, as well as a contribution from atmospheric transport  
77 were hypothesized (Vecchiato et al., 2017). Amongst the 17 fragrances, Benzyl, Hexyl and Amyl  
78 Salicylates were investigated by a limited literature: because of their low prices (under \$5/kg), these  
79 FMs are High Production Volume (HPV) chemicals with an worldwide annual consumption well over  
80 5000 tons (Belsito et al., 2007; Gaudin, 2014). Benzyl Salicylate is also an allergenic UV-filter agent  
81 (Heisterberg et al., 2011; Kameda et al., 2011), with an oestrogenic activity comparable to bisphenol  
82 A (BPA) (Zhang et al., 2012) and giving oestrogenic responses in human breast cancer cells (Charles  
83 and Darbre, 2009). Benzyl and Hexyl Salicylates were detected in different WWTPs (Godayol et al.,  
84 2015; Kameda et al., 2011; Negreira et al., 2010; Simonich et al., 2002, 2000), while fewer data are  
85 reported about Benzyl Salicylate in river waters (Kameda et al., 2011; Negreira et al., 2010; Vila et al.,  
86 2016).

87 More studies are available regarding Polycyclic Aromatic Hydrocarbons (PAHs) in the Mediterranean:  
88 the distribution of these compounds (deriving from incomplete combustions, direct dispersions of oil  
89 products and biogenic natural sources) was mainly investigated in coastal environments, while less  
90 studies considered the open sea areas (Berrojalbiz et al., 2011; Castro-Jiménez et al., 2012). The

91 available information shows that the occurrence of PAHs in offshore waters is dependent from  
92 atmospheric transport and deposition, particle and plankton settling processes in the water column  
93 and the presence of mesoscale variability associated with gyres (Berrojalbiz et al., 2011; Castro-  
94 Jiménez et al., 2012; Parinos and Gogou, 2016).

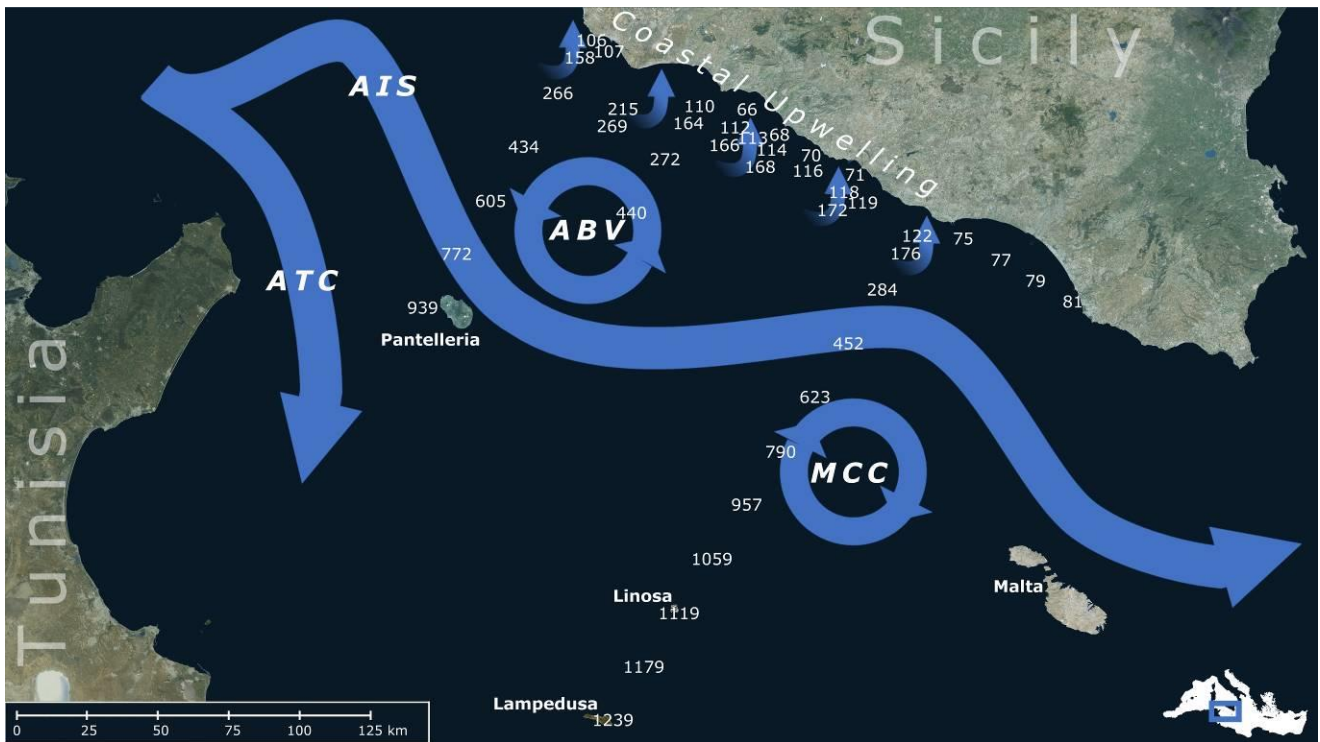
95 The aim of this work is to study the distribution of fragrances and PAHs in the surface seawater of  
96 the Mediterranean: the coupling of a new class of pollutants with well-known tracers, as PAHs,  
97 could help the interpretation of the processes governing their environmental presence. In  
98 particular, the detection of the selected FMs in open sea areas far from direct sources would  
99 highlight the role of the long-range atmospheric or oceanic transport, involving the possible  
100 environmental persistence of these contaminants. The study area is located in the Sicily Channel,  
101 which is the central connection between the Eastern and Western Mediterranean and the  
102 samplings were aimed to distinguish contamination in offshore and coastal areas. To the best of  
103 our knowledge this is the first survey investigating PCPs in offshore Mediterranean surface seawater  
104 and the first study focused on PAHs in the Sicily Channel, where only previous scarce data were  
105 available.

106

## 107 **2. Materials and methods**

108 Surface seawater samples (Table S12) were collected in July 2016 during the oceanographic cruise  
109 BANSIC 2016 (CNR R/V “Minerva Uno”) in solvent-rinsed glass bottles at the sites shown in Figure 1.

110



111

112 Figure 1: Sampling locations of the surface seawater and principal surface currents in the Sicily  
 113 Channel, Mediterranean. Modified from (Ciappa, 2009; Falcini et al., 2015).

114

115 In the Sicily Channel the surface Atlantic Water (AW) is advected by the Atlantic Ionian Stream (AIS)  
 116 and the Atlantic Tunisian Current (ATC) (Figure 1). In summer the meandering AIS flows round semi-  
 117 permanent structures: the cyclonic Adventure Bank Vortex (ABV), the anticyclonic Maltese Channel  
 118 Crest (MCC) (Ciappa, 2009; Falcini et al., 2015). An important spatial variability exists in terms of  
 119 shape, position and strength of these currents (Bonanno et al., 2014) and summer westerly winds  
 120 cause upwelling along the southern coast of Sicily (Ciappa, 2009; Patti et al., 2010). A deeper  
 121 westward current transports water of eastern origin at an intermediate layer (Bonanno et al., 2014)  
 122 and these hydrodynamic conditions also influence the distribution of biota in the area (Cuttitta et al.,  
 123 2017).

124

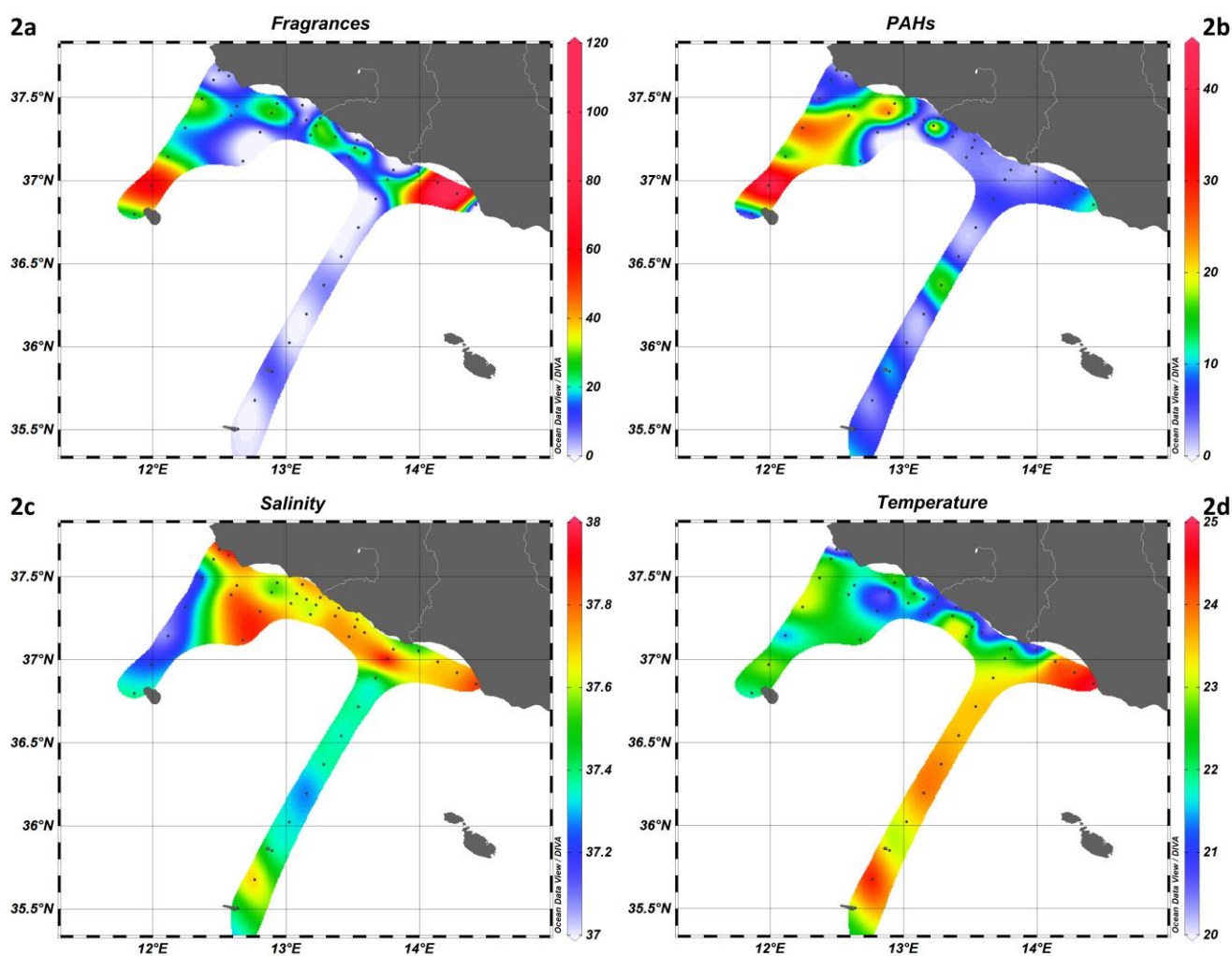
125 Details of the analytical method were previously described (Vecchiato et al., 2016, 2015). Briefly,  
126 water samples (1 L) were spiked with <sup>13</sup>C internal standards and extracted using 200 mg Oasis® HLB  
127 cartridges (Waters Corp., Milford, MA USA) previously conditioned with 10 mL of dichloromethane  
128 followed by 10 mL of ultrapure water (ELGA, High Wycombe, UK). Pesticide-grade dichloromethane,  
129 toluene, acetone and *n*-hexane (Romil Ltd., Cambridge, UK) were used. Isotope-labelled standard  
130 solutions (CLM-2477, CLM-2451, CLM-2722), were purchased from CIL (Cambridge Isotope  
131 Laboratories, Inc., Andover, MA USA), while PAHs native standard solutions (Naphtalene (NAP),  
132 acenaphthylene (ACY), acenaphthene (ACE), fluorene (FL), phenanthrene (PHE), anthracene (ANT),  
133 fluoranthene (FLT), pyrene (PYR), benzo(*a*)anthracene (B(*a*)A), chrysene (CHR), retene (RET),  
134 benzo(*b*)fluoranthene (B(*b*)F), benzo(*k*)fluoranthene (B(*k*)F), benzo(*a*)pyrene (B(*a*)P),  
135 benzo(*ghi*)perylene (B(*ghi*)P), indeno(1,2,3-*c,d*)pyrene (IcdP) and dibenzo(*a,h*)anthracene (D(*a,h*)A))  
136 were acquired from Dr. Ehrenstorfer (Augsburg, Germany). The results of the fragrances are  
137 expressed as a concentration of the commercial product standards (Givaudan, Vernier, Switzerland),  
138 since most of the 17 FMs are not available at analytical or reagent grade. Cartridges were eluted  
139 with 1 mL of toluene, 15 mL of dichloromethane followed by 10 ml of *n*-hexane. Eluates were dried  
140 with Na<sub>2</sub>SO<sub>4</sub> and reduced to 100 µL under a gentle nitrogen flow at 23 °C (Turbovap II®, Caliper Life  
141 Science, Hopkinton, MA, USA). Instrumental analyses were conducted by GC-MS (7890A-5975C,  
142 Agilent Technologies) on a 60-m HP-5MS column (0.25 mm I.D., 0.25 µm; Agilent Technologies,  
143 Avondale, USA) in Single Ion Monitoring (SIM). Crude results were corrected using the instrumental  
144 response factors and procedural field blanks (n=3). The Method Detection Limit (MLD) was  
145 calculated as three times the standard deviation of the blank signal and details are reported in Table  
146 S13.

147

148 **3. Results and Discussion**

149 **Fragrances**

150 Fragrances were detected in 27 out of 42 samples of surface seawater of the Sicily Channel, with the  
151 sum of the concentrations up to 112 ng L<sup>-1</sup> (Table SI4). Those compounds that were never detected  
152 in any sample (Amberketal, Dupical, Isobutavan, Lemonile Mefranal, Myraldene, Tridecene-2-Nitrile,  
153 Ultravanil) were excluded from the following discussion.



154

155 Figure 2a, b, c, d: Concentrations of Fragrances (ng L<sup>-1</sup>; 2a), PAHs (ng L<sup>-1</sup>; 2b), Salinity (PSU; 2c) and  
156 Temperature (°C; 2d) in seawater samples of the Sicily Channel. Refer to Figure 1 for sample stations  
157 (black dots) and to Tables SI4 and SI5 for other details. Data were interpolated using Ocean Data  
158 View - Diva Gridding.

160 Focusing on the western transect towards Pantelleria Island, the coastal samples resulted with lower  
161 levels of FMs in comparison to offshore waters, where concentrations relatively peaked in the  
162 sample 772 (64 ng L<sup>-1</sup>; Figure 2a) and in the near samples 605 and 939, resulting respectively 27 and  
163 26 ng L<sup>-1</sup>. A similar pattern marks also the coastal area between approximately 3 to 30 km from the  
164 Sicilian shore, where no or low FMs were found generally in correspondence to colder and saltier  
165 water masses (Figure 2c, 2d): concentrations ranging from 20 to 30 ng L<sup>-1</sup> were found at  
166 intermediate distance to the shore (samples 110, 164, 114, 116, 118, 119) in comparison to the  
167 samples 66, 70 and 71, resulting 1-10 ng L<sup>-1</sup> (Figure 3). Samples 68 and 113 except from this  
168 scheme. The different distribution between coastal and open sea areas is likely due to the  
169 occurrence of phenomena of wind-driven coastal upwelling raising deep and unpolluted waters  
170 (Ciappa, 2009; Patti et al., 2010), while the presence of gyres, as the cyclonic ABV, probably  
171 accumulates fragrances at their offshore borders, depending on the variability of the hydrodynamic  
172 patterns (Bonanno et al., 2014). However, the influence of the near Pantelleria island and/or the  
173 advection of AW from Northern Tunisian coastal areas cannot be ruled out. Different water masses,  
174 such as the anticyclonic MCC, influence the samples of the eastern transect towards Lampedusa  
175 Island, where, with few exceptions, fragrances resulted below detection limits. The highest  
176 concentrations of FMs were found in the samples 77 and 79 facing the gulf of Gela, resulting  
177 respectively 81 and 112 ng L<sup>-1</sup>. In this case it is possible to hypothesize the role of local land-based  
178 pollution, since in this area the upwelling is known to weaken (Piccioni et al., 1988), and marine  
179 currents transport eastwardly offshore the waters from the outlet of the Salso river, outflowing  
180 north-west to the gulf of Gela.

181 As in previous works (Vecchiato et al., 2017, 2016), Salicylates generally resulted the most abundant

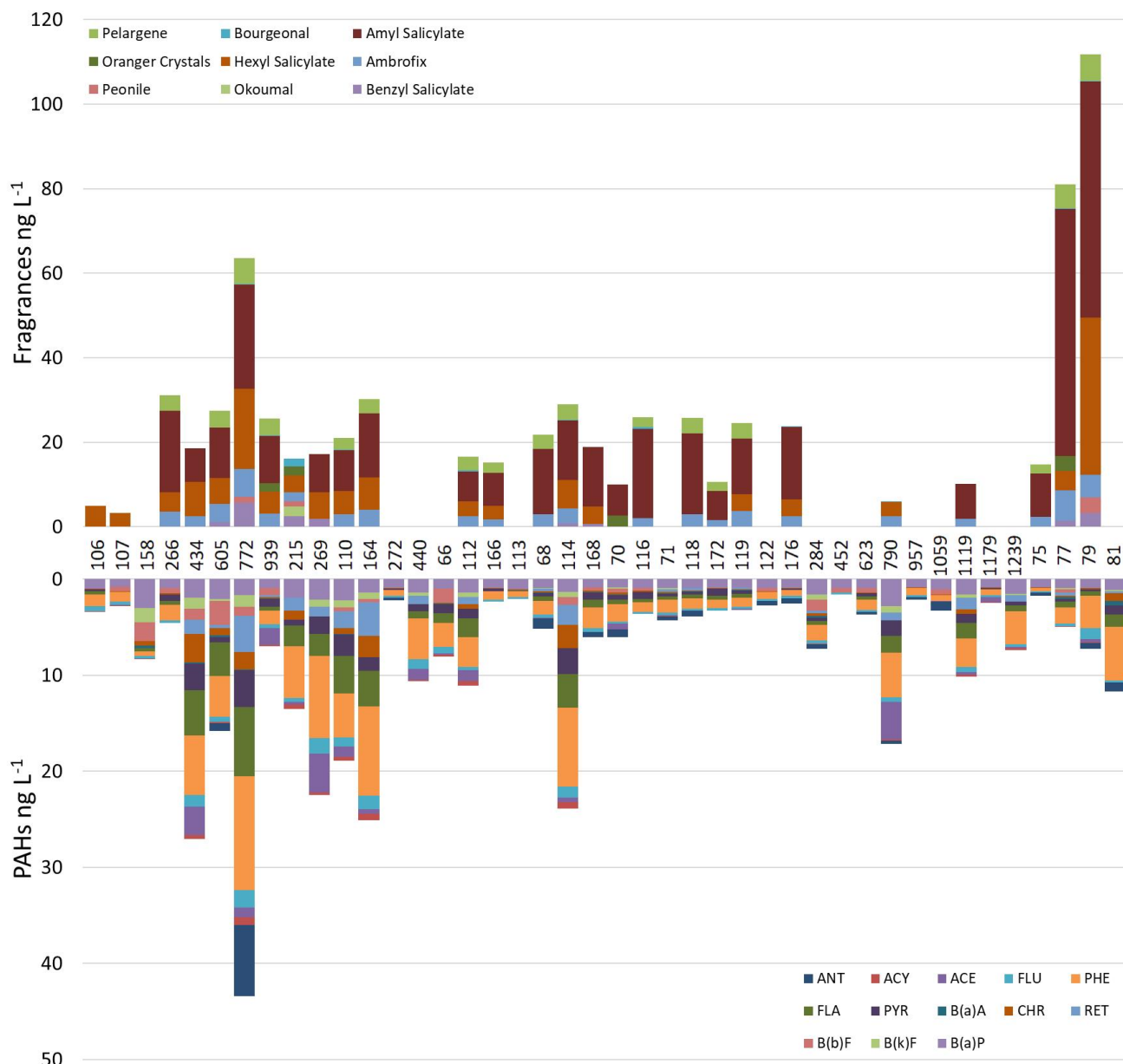
182 and frequently detected compounds, with the sum of Amyl and Hexyl Salicylate constituting the 25-  
183 100% of the total fragrances (Table SI4). Conversely, Benzyl Salicylate was detected only in 8 samples  
184 at 0.7-5.6 ng L<sup>-1</sup>: this differs with seawaters influenced by direct untreated discharges in the Venice  
185 lagoon (Vecchiato et al., 2016), but it is more similar to remote Antarctic samples, where  
186 atmospheric transport was hypothesized (Vecchiato et al., 2017). Higher levels of Benzyl Salicylate  
187 were found in wastewaters and rivers (Godayol et al., 2015; Kameda et al., 2011; Negreira et al.,  
188 2010; Simonich et al., 2002, 2000; Vila et al., 2016). The concentrations of this allergenic and  
189 oestrogenic compound (Charles and Darbre, 2009; Zhang et al., 2012) resulted far below possible  
190 ecotoxic damages to *Daphnia* and other species (Belsito et al., 2007).

191 Also Pelargene and Ambrofix were frequently detected in the samples, resulting respectively below  
192 6.3 and 7.2 ng L<sup>-1</sup>, while the occurrence of Bourgeonal, Oranger Crystals, Peonile and Okoumal was  
193 only occasional (Table SI4). In particular the last compound was found only in the sample 215, which  
194 was characterized by a peculiar composition, with amyl salicylate resulting below detection limits  
195 and a relatively higher proportion of the “minor” fragrances: it can be supposed that, considering the  
196 heavy maritime traffic in the area (Romagnoli et al., 2016), this specific sample reflects possible  
197 direct discharges from the previous passage of a ship.

198 The results of the FMs are in the same range of concentration of various PCPs globally detected in  
199 open sea environments (Arpin-Pont et al., 2016). Focusing on the Mediterranean region, the levels of  
200 FMs in the Sicily Channel are comparable to musk fragrances, UV-filters and other PCPs in  
201 environmental waters, resulting lower than samples from WWTPs and polluted hotspots (Cabeza et  
202 al., 2012; Celano et al., 2014; Gracia-Lor et al., 2012; Martínez Bueno et al., 2012; Negreira et al.,  
203 2010; Villa et al., 2012). The occurrence of PCPs also in biota (Cunha et al., 2015; Trabalón et al.,  
204 2015) and sediment matrices (Combi et al., 2016; Pintado-Herrera et al., 2016) accounts for the

205 ubiquitous presence of this class of compounds in the Mediterranean, with consequent concerns  
 206 about their environmental fate.

207



208

209 Figure 3: Fragrances and PAHs in the surface seawater of the Sicily Channel (ng L<sup>-1</sup>). Only detected  
 210 compounds are displayed.

211

## 212 PAHs

213 The presence of PAHs was detected in every sample of seawater, with concentrations ranging from  
214 1.6 to 43 ng L<sup>-1</sup> (Table S15). NAP, B(*ghi*)P, I(*cd*)P and D(*ah*)A resulted below detection limits and were  
215 not considered in the discussion. The distribution of PAHs is similar to fragrances (Figure 2b and 3):  
216 the highest levels were found in the sample 772 (43 ng L<sup>-1</sup>) along the transect towards Pantelleria,  
217 corresponding to the peak of FMs, followed by the sample 434 (27 ng L<sup>-1</sup>), while low concentrations  
218 were found near the coast, mainly due to upwelling phenomena. This is replicated at intermediate  
219 distance to the Sicilian shore, with peaks of concentrations in the samples 164 (25 ng L<sup>-1</sup>) and 114 (24  
220 ng L<sup>-1</sup>). Focusing on the eastern transect towards Lampedusa, relatively higher concentrations were  
221 found in the samples 790 (17 ng L<sup>-1</sup>) and 1119 (10 ng L<sup>-1</sup>), where also FMs resulted above detection  
222 limits. This pattern indicates that the processes of transport of both PAHs and FMs follow similar  
223 schemes and the presence of currents of water masses, as the ABV and the coastal upwelling, likely  
224 influences the distribution of these pollutants in the surface seawater. However, a remarkable  
225 exception is represented by the samples facing the gulf of Gela, where low levels of PAHs were found  
226 in the samples 77 and 79, contrary to the high concentrations of fragrances. This supports the  
227 hypothesis of a case of local pollution as a source of PCPs for these samples, in comparison to marine  
228 or atmospheric transport. Nevertheless, in Gela are located important petrochemical industries that  
229 might constitute a source of contamination to the surrounding environment, which is apparently in  
230 contrast with the relatively low levels of PAHs in our samples (Orecchio et al., 2010).

231 Concentration of PAHs in surface seawater are comparable to data from open sea areas of the  
232 Mediterranean (Berrojalbiz et al., 2011; Dachs et al., 1997; Ehrhardt and Petrick, 1993; Parinos and  
233 Gogou, 2016). In particular in the Sicily Channel, Dachs et al. found 1.1 ng L<sup>-1</sup> of PAHs, while  
234 Berrojalbiz et al. detected 0.25-0.76 ng L<sup>-1</sup> as a sum of dissolved and particulate fractions. Higher

235 levels (22.5 ng L<sup>-1</sup>) were found by Ehrhardt and Petrick as the mean concentration of transect in the  
236 same area (Berrojalbiz et al., 2011; Dachs et al., 1997; Ehrhardt and Petrick, 1993). In the coastal  
237 environments levels of  $\Sigma(\text{PAHs}) < 100 \text{ ng L}^{-1}$  are frequently found (El Nemr and Abd-Allah, 2003;  
238 Guigue et al., 2014; Guitart et al., 2010, 2004; Karacik et al., 2013; Lammel et al., 2015), however  
239 concentrations could rise substantially in presence of specific sources, river outlets and harbors  
240 (Bihari et al., 2007; Bouloubassi and Saliot, 1991; Er-Raioui et al., 2009; Marrucci et al., 2013;  
241 Montuori et al., 2016; Montuori and Triassi, 2012; Mzoughi and Chouba, 2011; Yilmaz et al., 1998).

242 Focusing on the individual compounds, FLU, PHE, FLA, PYR, CHR and B(a)P were detected in more  
243 than 70% of the samples and their sum constitutes the 53-99% of the total concentrations (Table  
244 SI5). PHE was the most abundant compound (up to 12 ng L<sup>-1</sup>), which is in agreement with previous  
245 data (Berrojalbiz et al., 2011). The average relative pattern of the major PAHs of the Sicily Channel is  
246 in accordance with that found in the seawater (Berrojalbiz et al., 2011) and in the aerosol (Castro-  
247 Jiménez et al., 2012; Tzapakis et al., 2006) collected in open sea areas, supporting the hypothesis of  
248 the role of the atmospheric transport leading net inputs of PAHs to open waters.

249 Retene (1-methyl-7-isopropyl phenanthrene) is a natural diterpenoid PAH deriving from the  
250 digenic degradation of abietic acid, a constituent of plant resins, but RET is also used as a  
251 molecular marker typical of combustion of coniferous wood (Bouloubassi and Saliot, 1993; Lipiatou  
252 and Saliot, 1991). The levels of RET in the surface seawater samples (l.d.l - 3.7 ng L<sup>-1</sup>) are comparable  
253 to previous data in the western Mediterranean (Bouloubassi and Saliot, 1993) and the occurrence of  
254 this compound suggests the role of biomass combustion as a major source of PAHs. This is also  
255 supported by the application of the diagnostic molecular ratios ANT/(ANT+PHE) and FLA/(FLA+ PYR)  
256 (Tobiszewski and Namieśnik, 2012), resulting in most samples respectively >0.1 and >0.5 (Figure  
257 SI1). These values indicate a clear prevalence of pyrogenic sources (mainly from combustion of

258 grass, wood and coal) upon petrogenic sources, which is in accordance with the distribution of PAH  
259 in the atmosphere over the open sea Mediterranean (Castro-Jiménez et al., 2012). Conversely  
260 petrogenic inputs are generally found to become more relevant in coastal areas (Gregoris et al.,  
261 2014). This suggests for PAHs a predominance of the atmospheric transport processes upon local  
262 pollution, e.g. fuel leaks from ships.

263

#### 264 **4. Conclusions**

265 This study represents the first detection of the selected FMs in open sea environment. The distance  
266 from direct sources evidences the previously hypothesized persistence of these contaminants  
267 (Vecchiato et al., 2017, 2016) involving processes of long-range transport. Both atmospheric and  
268 marine phenomena probably constitute key factors: on one hand the spatial distribution of the  
269 FMs, together with PAHs, can be explained considering the presence of hydrodynamic currents, as  
270 AW, ABV and coastal upwelling, on the other diagnostic ratios and literature data highlight the role  
271 of the atmospheric transport. With the exception of few coastal samples, the general concordance  
272 of the distribution of fragrances and PAHs indicates that both classes of compounds are depending  
273 from the same environmental drivers, as suggested also by the similarity of the range of their  
274 physico-chemical properties (Vecchiato et al., 2016). Consequently, it is likely the PAHs detected in  
275 the samples were originally emitted directly in the atmosphere by combustion processes, the FMs,  
276 as other PCPs, are generally supposed to enter in the environment through discharges in the water  
277 compartment (Vecchiato et al., 2016). Therefore, if an atmospheric transport is hypothesized, the  
278 volatilization probably occurs from wastewaters containing FMs, as well as directly from the  
279 household and personal goods where these products are used, since the semi-volatility is a  
280 fundamental commercial characteristic to give a persistent perfuming. Future studies should

281 disentangle the questions about the fate of the fragrances in the atmospheric compartment.  
282 Nevertheless, the peculiar concentrations of FMs found in the samples 77 and 79 plausibly reflect  
283 a local land-based pollution. The results highlight the role of mesoscale phenomena governing the  
284 distribution of organic pollutants between offshore and coastal areas, providing a first insight about  
285 the presence of fragrances in open seawater and improving the information available about the  
286 presence of PAHs in the Sicily Channel. This knowledge can help to achieve the challenging goal of  
287 the protection of the Mediterranean environment.

288

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297

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